INTRODUCTION

Coral reefs represent biogeographic centres of biodiversity and provide a large range of ecosystem goods and services (Moeburg & Folke 1999). Although we know that coral reefs are under increasing pressure from climate change, acidification, habitat destruction and overfishing (Floegh-Guldberg et al. 2007), we are still learning much about the factors controlling their productivity and function at an ecosystem level (Atkinson & Falter 2003). The relative importance of external and internal processes and nutrient sources for coral reef function has a long history of debate. Although it is now recognised that reefs may not be the self-sufficient ecosystems (Hatcher 1997) that early atoll studies had suggested (Sargent & Austin 1949, Odum...
that heterotrophic feeding by reef organisms may play an important role in the cycling of nutrients. Particulate nitrogen (N) is often considered to be a significant source of N, and the uptake of particulate N is confirmed to be effective under laboratory conditions (Wolanski & Delesalle 1995, Lesser et al. 2007) and internal wave activity (Ribes et al. 2003, Houlbrèque et al. 2004a). Such uptake of particulate N may be as high as that of dissolved N, indicating that particulate uptake is a potentially significant source of N. Several studies have confirmed that, under laboratory conditions, utilisation of particulate N may be as high as that of dissolved N (Ribes et al. 2003, Houlbrèque et al. 2004b) and that heterotrophic feeding by reef organisms may play a significant role in energy budgets, calcification (Ferrier-Pages et al. 2003, Houlbrèque et al. 2004a) and resilience to stressors (Grottoli et al. 2006, Palardy et al. 2008). Despite this, the dynamics of particles (including factors controlling their supply and incorporation into reef food webs) have received little attention, making the ecological significance of heterotrophy in coral reef systems, and by extension the biological oceanography of surrounding waters, especially difficult to quantify.

Quantifying particle fluxes in situ is often constrained by complex reef-scale habitat zonation and hydrodynamics (Crossland et al. 1991, Atkinson & Falter 2003). Attempts have been made to quantify particle flux in idealised systems, such as along channels (Fabricius & Dommasse 2000) or through perforated reefs (Fabricius et al. 1998); however, as Genin et al. (2009) point out, the flux inferred from these unique systems may be idiosyncratic. In a natural reef, variable flows and biogeochemical processes occur over different communities (e.g. Miyajima et al. 2007) and depths, in addition to variable lagoon residence times, vertical shearing of velocities and open lateral boundaries (Genin et al. 2009), can all confound attempts to calculate oceanic supply to, and particle removal by, a reef. As a result, a so-called control volume (CoVo) approach is often the most effective means of estimating the uptake of particulate nutrients by a reef (Genin et al. 2002, 2009). This approach depends on determining the flux of particles into and out of a defined control volume that contains the active uptake area (the reef), with any differences assumed to be due to uptake by the reef community. However, the CoVo approach does have restrictions: (1) reef hydrodynamics at the site must not be too complex (i.e. the upstream and downstream measurements occur at the ends of real streamlines); (2) overall the reef is viewed as a black box, eliminating the capacity for inferences regarding reef-scale variability in fluxes that are likely to occur among different reef zones (e.g. Miyajima et al. 2007); (3) the capacity for the simultaneous production and consumption of various components of suspended particulate matter within the CoVo are ignored (similar to simultaneous dissolved P uptake and remineralisation, Atkinson 1992), and particle uptake rates are therefore likely to be underestimated if reef production is not accounted for (Houlbrèque et al. 2006); and (4) a CoVo must be selected that is representative of the reef as a whole in terms of composition and processes if results are to be scaled beyond the CoVo.

In the present study, we examined phytoplankton fluxes across the fore reef, reef flat and lagoon of a reef section at Sandy Bay, Ningaloo Reef, on the northwest coast of Western Australia (22° 02' S, 113° 55' E; Fig. 1a), controlling for the factors that typically limit CoVo particle flux studies described above. The reef section is geometrically and hydrodynamically simple and well characterised, with sampling in different reef
Fig. 1. Sandy Bay, Ningaloo Reef, Western Australia. (a) Consistent shoreward cross-reef flow revealed by GPS drifters released shoreward of the surf zone (blue tracks) and outside the reef (yellow tracks); black arrows indicate direction of flow (see Fig. 2 for more details on hydrodynamics). (b) Reef zonation and bathymetry along black transect line in (a) derived from hyperspectral imagery. (c) Representative images of the benthic habitat demonstrate the shift from the tabulate hard coral dominated reef flat (Stn 3), to patch reef (Stn 4) and sandy lagoon along the flow path (Stn 5). Numbers 1 to 6 denote fixed water column sampling stations (see Fig. 4).
zones accounting for reef-scale variability in the benthic community. By focusing on phytoplankton, the influence of reef-level particle production on flux estimates is minimised, since the residence time over the reef is insufficient for pelagic phytoplankton productivity to significantly alter their concentrations, i.e. we were able to examine a predominately allochthonous particle source. Although the study focused on Sandy Bay, Ningaloo Reef as a whole is comprised of up to 50 functionally similar circulation cells along its length. In this system, consistent cross-reef flow resulting from wave-setup (Symonds et al. 1995, Lowe et al. 2009) in the surf zone drives oceanic water across the reef and into the lagoon. The reef–lagoon circulation cell thus represents a natural flume for quantifying particle flux in situ: this simple, well-defined natural flow is coupled with a natural reef community without the limitations associated with an experimental flume (e.g. limited dimensions, edge effects and limited forcing).

In addition to being a natural flume, the reef–lagoon circulation cells effectively link oceanographic- and reef-scale processes along the length of the reef. At ~290 km long, Ningaloo Reef is one of the largest fringing reefs in the world. The reef is close to shore, with a lagoon width as low as 200 m (average 2.5 km, Cassata & Collins 2008), but receives little to no terrestrial inputs from the adjacent arid desert other than infrequent storm discharge (Cassata & Collins 2008). The reef also lies very close to the edge of the continental shelf, which is at its narrowest along the Australian coast, and is thus impacted by 2 opposing coastal current systems that vary seasonally in relative strength: (1) the southward-flowing Leeuwin Current (LC), which is low nutrient and suppresses upwelling (Pearce 1991, Johannes et al. 1994), and (2) a northward-flowing (upwelling-favourable) Ningaloo Current (NC) that often dominates inshore of the LC in the summer months during periods of relatively strong southerly winds (Woo et al. 2006a), Ningaloo Reef thus experiences sporadic and transient upwelling during the austral summer. No direct comparison of oceanographic conditions in the 2 opposing seasons has been undertaken adjacent to the reef; however, sampling in LC and NC waters during the upwelling season has suggested that seasonal sporadic upwelling can increase nitrate concentrations 10-fold, from <0.1 to 2–6 µM, and more than double primary production, from 100–500 mg C m² d⁻¹ to 850–1300 mg C m² d⁻¹ (Hanson et al. 2005). This backdrop of dynamic regional oceanography means that Ningaloo Reef is an ideal location for examining the role of external oceanographic processes in reef biogeochemistry. While the reef–lagoon section at Sandy Bay exhibits relatively consistent circulation patterns during the year, nutrient loading (both dissolved and particulate) is thought to vary significantly among seasons and perhaps over relatively short time scales (e.g. days to weeks) in response to oceanographic forcing (Feng & Wild-Allen 2009). In the present study we examined the significance of reef-level phytoplankton flux at Ningaloo Reef in the context of regional oceanographic processes.

MATERIALS AND METHODS

Study site. Sandy Bay, Ningaloo Reef, is located near the centre of the North West Cape (Fig. 1a) and is the site of the most comprehensive reef hydrodynamic studies conducted on Ningaloo (Hearn & Parker 1988, Lowe et al. 2008). The reef morphology is characterised by a ~1.30 fore reef slope that rises to a shallow (~1.5 m) reef flat stretching ~500 m shoreward of the surf zone (Fig. 1b). The reef flat is separated from shore by a ~500 m wide, ~2 to 3 m deep lagoon. Surface waves of ~1 to 3 m are incident to Ningaloo throughout the year; wave-setup generated by these breaking waves in the surf zone (Symonds et al. 1995, Lowe et al. 2009) elevates the mean water level in the vicinity of the reef crest, establishing a pressure gradient that drives a cross-reef flow. Water driven into Ningaloo’s lagoons by wave-setup returns back to the ocean through gaps (channels) in the reef (Fig. 1a). The mean tidal amplitude at Ningaloo is relatively small (~0.5 m), such that at Sandy Bay the circulation is dominated by wave forcing, i.e. tides modulate the flow across the reef but it very rarely reverses (see ‘Results: hydrodynamics’).

The reef composition at Sandy Bay is typical of that observed elsewhere on Ningaloo Reef. In general terms, the reef is characterised by a poorly studied fore reef environment rising to a reef crest with high cover of encrusting coralline algae (~80%, Cassata & Collins 2008), a reef flat with high cover of tabulate scleractinian corals (Acropora spp., often >100% cover) and large rubble and sand patches, a patchy back reef habitat including large stands of branching staghorn Acropora spp. and a sandy lagoonal habitat interspersed with algal patches. Visual surveys of percentage cover within 1 m² quadrats undertaken in May 2007 revealed that the dominant average cover (>10%) in the reef flat zone at Sandy Bay was contributed by rubble (24%), sand (23%) and tabulate Acropora spp. coral (18%) (A. S. J. Wyatt unpubl. data). High (~10%) but patchy cover by bottlebrush Acropora spp. and soft corals was also evident. The lagoon was 58% sand and 26% algae (largely Sargassum sp.). The fore reef composition at Ningaloo is generally poorly known (Cassata & Collins 2008). Quantification of the reef slope community via quadrats was not undertaken directly in the present study, as it is
logistically difficult by SCUBA diving due to water depths, strong currents and large swells. From previous studies, the reef slope to about 30 m depth is characterised by spur-and-groove formations supporting a diverse hard coral assemblage dominated by tabulate Acropora spp., with soft coral, Millepora, sponges and macroalgae (Cassata & Collins 2008). Visual observations around Stns 1 and 2 (20 and 15 m depth, respectively; Fig. 1a,b) suggest increasing dominance by Porifera (sponges) and Alcyonaria (sea whips and gorgonians) with increasing depth (Colquhoun et al. 2007), but scleractinian corals (Acropora spp.) continued to dominate live cover in this depth range.

Field experiment. Four separate field experiments of ~3 to 6 wk duration were conducted in both May and November 2007 and 2008. During each experiment, Eulerian water column sampling was regularly conducted at 6 fixed stations along the dominant cross-reef flow path (see below and Fig. 1a), the exception being November 2007 when fixed sampling was only possible at Stns 3 and 4. The reef slope stations (Stns 1 and 2) were located so as to be as distant as possible from the potential influence of channel outflow. Lagrangian sampling was also undertaken in November 2007 and May 2008 along flow paths (at ~20 min intervals) taken by 2 GPS-tracked cruciform drifters (see ‘Hydrodynamics’; Austin & Atkinson 2004), adjacent to the drifters as they traversed the reef.

Due to time required to collect and analyse water samples, it was generally not possible to sample all stations each day; instead, a portion of the stations were typically sampled on alternate days. The exception was during the November 2008 experiment where most stations (i.e. Stns 1 to 4) were sampled simultaneously on most days.

Hydrodynamics. During each experiment, mean currents within the circulation cell were monitored continuously using a moored 2 MHz Nortek Aquadopp profiler (ADP) deployed on the reef flat between Stns 3 and 4. Hourly depth-averaged current speeds $U_h$ across the reef were computed from raw 3-dimensional velocity profiles (0.1 m bin size) recorded every 15 min. Time series of the local water depth $h_t$ inferred from the ADP pressure sensor were used to calculate a volumetric flow rate per unit reef width $q_v = U_h h_t$.

In addition to moored current measurements, Lagrangian observations of water circulation were made through the periodic deployment of 2 drifters (see above). The drifters were initially released ~50 m shoreward of the surf zone and typically followed until they reached the channel entrance (Fig. 1a). A limited number of deployments were also possible outside the reef (near Stn 2), with the drifters allowed to travel through the surf zone and over the reef (Fig. 1a). The drifters extended ~0.5 m below the water surface and were thus just able to traverse the reef flat during spring low tides.

In May and November 2008, a 1 MHz Nortek AWAC current profiler/directional wave gauge was deployed on the fore reef in ~20 m of water near Stn 1 (Fig. 1a). Hourly wave spectra were computed using raw time series of water surface variability collected using acoustic surface tracking. These spectra were used to compute time series of significant wave height $H_s$ and peak period $T_p$ for the waves incident to the reef.

Sampling and analysis. At each sampling location, 2 replicate 20 l surface water samples were taken and filtered within 2 h for chl $a$ and particulate nutrient concentrations. For total chl $a$ analysis, 11 samples were filtered onto Whatman GF/F filters and stored frozen until overnight extraction in 8 ml 90% acetone within 10 d of collection. For analysis of the $>5$ µm fraction, 2 l of seawater were filtered onto 5 µm Nitex mesh. All samples were analysed for chl $a$ with a fluorometer (Turner Designs T700) without grinding. Samples were acidified with 10% HCl to correct for phaeopigments, and chl $a$ concentrations were calculated according to Parsons et al. (1984). All analyses were performed using total chl $a$ data except where explicitly stated (i.e. the small to large cell ratio analysis).

POC and PON values were obtained by filtering 4 l of water through precombusted (500°C, 6 h) Whatman GF/F filters. Samples were acidified to remove carbonates using either acid fumigation or direct drop-wise addition of 1 M HCl. Samples were stored frozen and desiccated until analysis. Samples were combusted in tin capsules using an elemental analyser (ANCA-GSL, Europa Scientific) to N$_2$ and CO$_2$. N$_2$ and CO$_2$ were purified by gas chromatography and the N and C elemental composition determined by continuous flow-isotope ratio mass spectrometry (20-20 IRMS, Europa Scientific).

It has been found that acid fumigation is insufficient to eliminate erroneously high POC values resulting from carbonate contamination in shallow-water coral reef samples due to significant sand resuspension (A. S. J. Wyatt et al. unpubl. data). Consequently, POC data obtained using acid fumigation and direct addition of HCl showed varying levels of carbonate sand contamination. Samples with $\delta^{13}$C values above -9‰ were considered to contain carbonate sand and were excluded from POC analysis. Both methods of acidification were found to have no significant effect on PON (A. S. J. Wyatt et al. unpubl. data).

Phytoplankton C and N (POC$_p$ and PON$_p$, respectively) were estimated by correcting for non-photo-synthetic components of total POC and PON using the same correction conventionally applied to POC (Banse 1977).

Phytoplankton uptake rates. Across the reef (from Stns 1 to 4), where the cross-reef flow was approxi-
mately 1-dimensional (see below), the volumetric flow rate \( q \) (not the current speed \( U \)) should be roughly conserved (Lowe et al. 2009). This was supported by paired drifter releases along the study transect (e.g. Fig. 1), which were not observed to significantly diverge or disperse; given that these drifter paths represent fluid streamlines, the volume of flow between adjacent streamlines would be roughly conserved (Kundu & Cohen 2002). From the measured chl a concentration \( C_i \) at each station \( i \), the phytoplankton uptake rate \( m_i \) (mg m\(^{-2}\) d\(^{-1}\)) between each pair of sampling stations can thus be estimated as:

\[
m_i = q \frac{\Delta C_i}{\Delta x_i}
\]

where \( \Delta C_i = C_i - C_{i+1} \) is the concentration difference between adjacent sites and \( \Delta x_i = |x_i - x_{i+1}| \) is the distance between stations. Note that positive uptake values correspond to a removal of phytoplankton. In practice, given that it was generally not possible to sample all sites on a single day, \( m \) was calculated using the average chl a profile across the reef over consecutive 2 d periods and the average flow rate \( q \) during this time.

To investigate spatial variability in \( m \), values were calculated for 4 distinct regions: (1) fore reef (Stns 1 to 2), (2) reef crest including the surf zone (Stns 2 to 3), (3) reef flat (Stns 3 to 4), and (4) lagoon (average of Stns 4 to 5 and 5 to 6). Finally, for the case where particle uptake by the benthos is physically controlled (i.e. mass transfer limited, Sanford & Crawford 2000; Ribes et al. 2003), uptake is often related to the water column concentration \( C \) as \( m = SC \), where \( S \) is a particle uptake coefficient (m s\(^{-1}\)).

**Statistical analysis.** Statistical analysis was performed in SPSS v17.0. A general linear model ANOVA was used to check for significant differences between groups after confirming homogeneity of variance using Levene’s test. Post hoc differences were examined with Fisher’s LSD. Proportional data were arcsine transformed.

**RESULTS**

**Hydrodynamics**

During May 2008, significant wave heights \( H_s \) ranged from 0.8 to 2.5 m (Fig. 2b) and varied between 1 and 3 m during May 2007 and November 2008. For all experiments, the depth-averaged current speed across the reef flat \( U_f \) (between Stns 3 and 4) varied from ~10 to 30 cm s\(^{-1}\) (Fig. 2c). Moreover, currents at this location were oriented directly across the reef (deviating on average <10° from the shore-normal direction) and never reversed (Fig. 2d), including during periods with ebbing spring tides (Fig. 2a).

Water therefore roughly flows along a transect from Stns 1 to 6 (total transect length = 3000 m). Given the typical current speeds in this circulation cell (~10 to 30 cm\(^{-1}\)), the residence time of water in the reef--
Wyatt et al.: Coral reef phytoplankton fluxes

lagoon system (from entering the reef to exiting the channel) would be ~2 to 6 h. Wave breaking in the surf zone is clearly the dominant hydrodynamic forcing mechanism driving the circulation, with the cross-reef current speed $U_r$ increasing roughly linearly with the incident wave height $H_s$ (Fig. 3). There was a strong correlation between $U_r$ and $H_s$ observed for all experiments ($r^2 = 0.64$, $F_{353} = 627.2$, $p < 0.001$), indicating that a reasonable estimate of the currents can also be made from measurements of $H_s$ alone in this predominantly wave-driven flow environment.

A total of 29 paired drifter releases across the 4 experimental periods confirmed the current meter observations and revealed a persistent and steady circulation pattern in the reef–lagoon system, with a shoreward flow across the reef flat and a return flow out the channel (Fig. 1a). On one occasion the drifters were allowed to cross the reef crest through the surf zone and move across the reef (Fig. 1a). This flow pattern was evident under a range of wind conditions with little deviation even under strong winds (e.g. ~13 m s$^{-1}$).

Phytoplankton: spatial and temporal variability

Phytoplankton concentrations were markedly depleted as water flowed over the reef. During each field season the mean chl $a$ concentration profile along the transect (Fig. 4a) showed a decline of 50 to 60% as water moved from the fore reef (Stn 1) to the back reef (Stn 4), yet remained relatively constant through the lagoon (Stns 4 to 6). In November 2007, Eulerian sampling was only possible at Stns 3 and 4, but as shown in Fig. 4, chl $a$ concentrations were similar at both stations to corresponding values in November 2008. Lagrangian observations demonstrated similar patterns of chl $a$ depletion as water moved over the reef flat, with associated uptake rates in good agreement with Eulerian estimates on the reef flat (see below and Table 3). Concomitant with the decreases in chl $a$ over the reef flat was an increase in phaeophytin concentrations (Fig. 4b), indicative of chlorophyll breakdown (reviewed by Jeffery et al. 1997). Phaeophytin concentrations followed a pattern approximating the inverse

Table 1. Number of replicate water samples (2 × 20 l) collected during each of the 4 field experiments for Eulerian (totals and samples per station) and Lagrangian sampling

<table>
<thead>
<tr>
<th>Experiment</th>
<th>Total</th>
<th>Eulerian</th>
<th>Lagrangian</th>
</tr>
</thead>
<tbody>
<tr>
<td>May 2007</td>
<td>46</td>
<td>8</td>
<td>8</td>
</tr>
<tr>
<td>Nov 2007</td>
<td>8</td>
<td>0</td>
<td>0</td>
</tr>
<tr>
<td>May 2008</td>
<td>27</td>
<td>5</td>
<td>6</td>
</tr>
<tr>
<td>Nov 2008</td>
<td>41</td>
<td>8</td>
<td>9</td>
</tr>
<tr>
<td>Total</td>
<td>122</td>
<td>21</td>
<td>22</td>
</tr>
</tbody>
</table>

Fig. 3. Significant relationship between cross-reef current velocities ($U_r$, cm s$^{-1}$, positive towards shore) measured on the reef flat and the significant wave height ($H_s$, m) measured on the fore reef ($r^2 = 0.64$, $F_{353} = 627.2$, $p < 0.001$). Line represents the best least-squares fit: $U_r = 6.5H_s + 5.0$

Fig. 4. Eulerian sampling showing (a) chlorophyll $a$ (chl $a$) and (b) phaeophytin concentrations in water flowing over Ningaloo Reef in May and November 2007 and 2008. Numbers: sampling stations as in Fig. 1a. Means ± SE, n as per Table 1
of chl a, with phaeophytin increasing by a factor of ~3 to 5 at Stn 4, after which it declined in the lagoon.

Phytoplankton cells were predominately small, with an average of ~83% of total chlorophyll in oceanic waters (Stns 1 and 2) contributed by cells <5 µm (Table 2). There was a significant difference in the arcsine-transformed proportion of small phytoplankton between the full field seasons (May 2007, May and November 2008), due to a lower proportion of small cells in May 2008 ($F_{2,94} = 8.599$, $p < 0.001$), particularly inside the reef. The proportion of small plankton remained relatively constant within the reef (Stns 4 to 6: ~79% in May 2007, 60% in May 2008 and 87% in November 2008) and there were no significant differences between reef zones or a significant interaction term between season and zones in the 2-way ANOVA model.

Marked temporal variability in phytoplankton supplied to the reef was evident at a range of temporal scales. At the seasonal scale, chl a concentrations outside the reef (Stn 1) were very similar between both May experiments (~0.7 µg l$^{-1}$), but were much lower during the November 2008 experiment (~0.4 µg l$^{-1}$). At the time scale of individual experiments (weekly) variation was also evident; during the periods when sampling outside the reef was possible, chl a concentrations varied markedly in time by a factor of ~5 (typically ranging from 0.2 to 1.0 µg l$^{-1}$; Fig. 5). Although the corresponding lagoon chl a levels increased with offshore concentrations, the variation in lagoonal concentrations during experiments was relatively small (~0.1 to 0.3 µg l$^{-1}$). Marked temporal variability sometimes also occurred over a time scale of just days, as evident during sampling in November 2008 (Fig. 6a). The initially high chl a levels outside the reef (~0.6 µg l$^{-1}$) dropped in a matter of only a few days to ~0.2 µg l$^{-1}$, then gradually recovered to ~0.6 µg l$^{-1}$ during the final week of the experiment. There was some evidence that these declines were related to the influence of the LC, with lower concentrations aligning with a shift to warmer (Fig. 6c) flow to the south (Fig. 6b) at Stn 1 (Fig. 1a). Conversely, the northward, colder flow prevailing during most of November 2008 is indicative of NC water (Fig. 6b,c).

### Phytoplankton uptake

The depletion of phytoplankton was pronounced during passage over the reef crest and the concomitant increase in chl a breakdown (phaeophytin concentrations; Fig. 4b) suggests that chl a was being consumed. The time scale of passage over the reef (~2 to 6 h total, <1 h over the active reef area) suggests that phytoplankton cell growth and death processes are unlikely factors in the changes in chl a observed (e.g. Fabricius & Dommisse 2000).

Phytoplankton uptake rates $m$ were calculated for each of the 4 regions (Eq. 1). Average phytoplankton uptake rates for all experiments revealed that uptake only occurred over regions in the system covered by coral communities (Table 3): ~8 mg m$^{-2}$ d$^{-1}$ for the fore reef, ~4 mg m$^{-2}$ d$^{-1}$ for the reef crest, ~6 mg m$^{-2}$ d$^{-1}$ for the reef flat and ~0 mg m$^{-2}$ d$^{-1}$ within the lagoon. Lagrangian observations were in very close agreement with these uptake rates, indicating that the LC was driving phytoplankton depletion in the reef system.

### Table 2. Phytoplankton (<5 µm) during each of the 4 field experiments and at each sample station. Means ± SE %, n as per Table 1. ‘–’ indicates no data available

<table>
<thead>
<tr>
<th>Experiment</th>
<th>Total</th>
<th>Station</th>
</tr>
</thead>
<tbody>
<tr>
<td></td>
<td></td>
<td>1</td>
</tr>
<tr>
<td>May 2007</td>
<td>80.84 ± 2</td>
<td>82.2 ± 8</td>
</tr>
<tr>
<td>Nov 2007</td>
<td>62.96 ± 5</td>
<td>–</td>
</tr>
<tr>
<td>May 2008</td>
<td>68.62 ± 5</td>
<td>78.6 ± 14</td>
</tr>
<tr>
<td>Nov 2008</td>
<td>87.66 ± 1</td>
<td>89.9 ± 3</td>
</tr>
<tr>
<td>Total</td>
<td>83.7 ± 5</td>
<td>83.6 ± 5</td>
</tr>
</tbody>
</table>
with uptake rates obtained from Eulerian measurements over the reef flat (Table 3). The pattern of decreasing uptake across the reef was generally consistent between experiments (Fig. 7a), although in May 2007 there was highly variable uptake at the reef crest leading to a significant interaction between experiment and reef zone terms in the ANOVA model ($F_{6,124} = 2.693$, $p = 0.018$). Consequently, experiments were analysed independently. Fore reef uptake rates were significantly higher than those on the reef flat during May and November 2008 and those in the lagoon for all 3 experiments (May 2007: $F_{3,54} = 6.829$, $p = 0.001$; May 2008: $F_{3,29} = 15.40$, $p < 0.001$; November 2008: $F_{3,41} = 13.01$, $p < 0.001$). Reef crest uptake was also significantly higher than in the lagoon during May 2008. Analysis of individual zones revealed no significant differences in uptake rates between experiments except for lower uptake rates on the reef flat during November 2008 ($F_{2,35} = 11.36$, $p < 0.001$).

To evaluate the total phytoplankton removal rate $M$ by the 3 reef-covered sections (per unit of alongshore reef width, excluding the lagoon where uptake was negligible), $m$ was spatially integrated according to:

$$M = \sum_{i=1}^{3} (m_i \Delta x_i)$$

This gives $M$-values of 5000, 7600 and 2900 mg m$^{-1}$ d$^{-1}$, for the May 2007, May 2008 and November 2008 experiments, respectively.

Given the temporal variability in chl a profiles in Fig. 6a, the total rate of particle removal $M$ also has the potential to vary on a time scale of days to weeks. Using the 5 profiles from Fig. 6a and Eqs. (1) & (2), the temporal variability in $M$ during the November 2008 sampling period can be calculated (Fig. 8). Initially on 7 to 8 November, the relatively high chl a concentrations outside the reef led to a value of $M = 7900$ mg m$^{-1}$ d$^{-1}$, which is comparable to the average May 2007 rate.

Particle uptake coefficients $S$, computed as $S = m/C$, where the concentration is taken as the mean value

---

Table 3. Phytoplankton uptake rate $m$ (mg chl a m$^{-2}$ d$^{-1}$), uptake coefficient $S$ (m d$^{-1}$) and corresponding phytoplankton particulate organic nitrogen (PON$_p$) and carbon (POC$_p$) flux (mmol m$^{-2}$ d$^{-1}$) determined in each reef zone and by Lagrangian sampling over the reef flat. Means ± SE, n as per Table 1. Dimensionless Stanton numbers ($\times 10^{-3}$) are also shown (see ‘Discussion’).

<table>
<thead>
<tr>
<th>Reef zone</th>
<th>$m$</th>
<th>$S$</th>
<th>PON$_p$</th>
<th>POC$_p$</th>
<th>$St$</th>
</tr>
</thead>
<tbody>
<tr>
<td>Fore reef</td>
<td>8.03 ± 0.79</td>
<td>19.2 ± 2.0</td>
<td>4.87 ± 0.48</td>
<td>27.9 ± 2.7</td>
<td>1.48 ± 0.15</td>
</tr>
<tr>
<td>Reef crest</td>
<td>4.10 ± 1.40</td>
<td>9.33 ± 3.9</td>
<td>2.48 ± 0.85</td>
<td>14.2 ± 4.8</td>
<td>0.720 ± 0.30</td>
</tr>
<tr>
<td>Reef flat</td>
<td>5.67 ± 0.77</td>
<td>21.3 ± 2.8</td>
<td>3.43 ± 0.47</td>
<td>19.6 ± 2.7</td>
<td>1.64 ± 0.22</td>
</tr>
<tr>
<td>Lagoon</td>
<td>−0.110 ± 0.25</td>
<td>−0.278 ± 1.1</td>
<td>−0.067 ± 0.15</td>
<td>−0.383 ± 0.87</td>
<td>−0.021 ± 0.08</td>
</tr>
<tr>
<td>Lagrangian</td>
<td>5.49 ± 1.1</td>
<td>20.2 ± 4.6</td>
<td>2.72 ± 0.68</td>
<td>15.6 ± 3.9</td>
<td>1.56 ± 0.35</td>
</tr>
</tbody>
</table>
As for \( m \), there was a significant interaction between experiment and reef zone (\( F_{6,114} = 3.304, p = 0.005 \)). In general, uptake coefficients were highest on the fore reef and reef flat, averaging ~19 and 21 m d\(^{-1}\), respectively (Table 3). During all experiments, \( S \) was negligible in the lagoon. In May 2007, uptake coefficients were significantly higher on the fore reef and reef flat (\( F_{3,52} = 7.903, p < 0.001 \)). In May 2008, they were lower (negligible) in the lagoon (\( F_{3,25} = 10.62, p < 0.001 \)) with no other differences between zones. In November 2008, \( S \) was significantly higher on the fore reef relative to all other zones and also on the reef flat relative to the lagoon (\( F_{3,37} = 13.95, p < 0.001 \)). Analysing by reef zone, there were no significant differences between experiments, apart from lower uptake coefficients on the reef flat during November 2008 (\( F_{2,33} = 8.190, p = 0.001 \)).

between the adjacent sections, are shown in Fig. 7b.

Phytoplankton nutrient uptake (POC and PON)

In order to estimate nutrient fluxes due to the phytoplankton uptake observed, chl \( a \) concentrations were correlated with direct measurements of POC and PON. Cross-reef particulate production complicates direct calculation of POC and PON uptake and examination of only the phytoplankton proportion of nutrient flux is an effective proxy for examining allocthonous (ocean-derived, as opposed to reef-produced) particulate nutrients.

There was a significant relationship between PON and chl \( a \) \( (r^2 = 0.305, F_{1,98} = 42.95, p < 0.001) \) that revealed a PON:chl \( a \) ratio of 8.48 ± 1.3. There was no significant difference in the PON–chl \( a \) relationship between May 2007, May 2008 and November 2008 (ANCOVA interaction term between chl \( a \) and season, \( F_{2,94} = 2.474, p = 0.09 \)). A large proportion (~30%) of the variation observed in PON could be attributed to variation in phytoplankton concentrations (chl \( a \)), with detritus contributing 6.03 ± 0.53 µg N l\(^{-1}\) of the total PON. This relationship allows phytoplankton PON (PON\(_p\)) to be estimated by:

\[
PON_p = 84.8C
\] (3)

where \( C \) is the chl \( a \) concentration. Using the average phytoplankton uptake rates (\( m \)) for each reef zone, the equivalent PON\(_p\) flux was then calculated (Table 3). The highest flux was evident over the reef front, equivalent to ~5 mmol N m\(^{-2}\) d\(^{-1}\), relative to ~2 and 3 mmol N m\(^{-2}\) d\(^{-1}\) over the reef crest and reef flat, respectively, with negligible flux in the lagoon. The spatially integrated phytoplankton uptake (\( M \)) determined using Eq. (2) and converted to PON\(_p\) using Eq. (3) gives PON\(_p\) = 42 400, 64 448 and 24 592 mg N m\(^{-1}\) d\(^{-1}\), for the
May 2007, May 2008 and November 2008 experiments, respectively.

Although there was a significant relationship between phytoplankton and POC, chl \(\alpha\) concentrations only explained \(\sim 12\%\) of the variation in POC \((r^2 = 0.117, F_{1,88} = 11.66, p < 0.01)\). As for PON, there was no significant seasonal difference in the POC–chl \(\alpha\) relationship \((\text{ANCOVA interaction term between chl} \alpha \text{ and season}, F_{2,84} = 0.7284, p = 0.49)\). The contribution of detritus to POC across all stations in May 2007, May 2008 and November 2008 was \(54.1 \pm 4.9 \mu g \text{ C l}^{-1}\).

Phytoplankton POC \((\text{POC}_{p})\) can be estimated as:

\[
\text{POC}_{p} = 41.6C
\]

where \(C\) is the chl \(\alpha\) concentration, leading to an equivalent \(\text{POC}_{p}\) flux for each reef zone \((\text{Table 3})\): \(\sim 28 \text{ mmol} C \text{ m}^{-2} \text{ d}^{-1}\) on the fore reef compared to \(\sim 14\) and \(\sim 19 \text{ mmol} C \text{ m}^{-2} \text{ d}^{-1}\) on the reef crest and flat, respectively. POC flux was negligible inside the lagoon. The spatially integrated phytoplankton uptake \((M)\) converted to \(\text{POC}_{p}\) using Eq. \((4)\) gives \(\text{POC}_{p} = 208,316\) and \(121 \text{ g} C \text{ m}^{-2} \text{ d}^{-1}\) for the May 2007, May 2008 and November 2008 experiments, respectively.

**DISCUSSION**

Significant supply and assimilation of externally produced particulate matter has a number of implications for our understanding of the function of coral reefs in terms of nutrient dynamics and the influence of oceanographic processes. The simple wave-driven cross-reef circulation documented in the present study at Ningaloo suggests that it is an ideal system for quantifying this uptake of ocean-derived material, while the marked seasonal changes in the adjacent ocean make it possible to begin assessing the influence of oceanographic processes on reef-level biogeochemistry.

**Phytoplankton fluxes**

The supply of oceanic material to the shallow Ningaloo Reef community is largely wave-driven, with little influence from tide and wind forcing. Based on average flow conditions, wave-driven flow typically leads to a volume flux of \(\sim 19,000 \text{ m}^3 \text{ d}^{-1}\) across this reef flat per meter length of coastline; for the entire length of Ningaloo \((290 \text{ km}, 85\% \text{ reef})\) this equates to \(\sim 4.79 \times 10^6 \text{ m}^3\) of ocean water crossing the reef into the lagoon every day.

The observed phytoplankton uptake within these wave-driven flows was clearly linked to the benthic reef community. Phytoplankton were removed from the water column only over the reef zones, at an average rate for \(m\) of \(\sim 4\) to \(8 \text{ mg} \text{ chl} \alpha \text{ m}^{-2} \text{ d}^{-1}\), with negligible removal within the sandy lagoon. Lagrangian phytoplankton uptake observations were in close agreement with Eulerian measurements, reinforcing the simple nature of the hydrodynamics that facilitates using a CoVo approach with fixed sampling stations. As observed in previous studies \((\text{Fabricius \\& Dommisse 2000, Houlbrèque et al. 2006})\), the residence time over reefs \((\sim 2\) to \(6 \text{ h}\) total, \(< 1 \text{ h}\) over the active reef) is insufficient to explain differences in phytoplankton between stations based on cell growth rates. Phytoplankton growth rates over the reef based on ammonium uptake rates of \(\sim 3 \text{ nmol} \text{ l}^{-1} \text{ h}^{-1}\) \((\text{Kapeli 2007})\) suggest a doubling time of \(> 60 \text{ h}\), which clearly places phytoplankton growth as insignificant at the time scale of the cross-reef flow. Regardless, significant phytoplankton growth between stations would actually imply even greater particle uptake than we calculated from the declining concentrations between stations.

The specific mechanisms for phytoplankton depletions, such as direct consumption by filter feeders or less direct processes, such as physical trapping in the reef framework and microbial degradation, requires further examination. While purely physical removal based on settling of cells as water moves across the reef is possible, we would expect the greatest settlement to occur in the less energetic back reef and lagoon zones — the observed flux rates were actually lowest in these zones. Furthermore, uptake of chl \(\alpha\) was comcomitant with an increase in phaeophytin, which is indicative of phytoplankton breakdown and argues against purely physical removal. Whether phytoplankton were degraded due to direct consumption by filter feeders \((\text{i.e. grazing})\), or trapped and then degraded \((\text{e.g. in coral mucus, Naumann et al. 2009})\), it is clear that there is a flux of ocean-derived phytoplankton to the reef which are, presumably, assimilated into the reef food web \((\text{i.e. grazing and physical trapping are likely to be biogeochemically equivalent, Genin et al. 2009})\).

Phytoplankton fluxes were somewhat larger on the fore reef where phytoplankton concentrations were highest. Uptake coefficients, however, suggest as great a potential for uptake by the reef flat community \((S = 19 \text{ m} \text{ d}^{-1})\) as the fore reef \((S = 21 \text{ m} \text{ d}^{-1})\). Higher rates of hydrodynamic energy dissipation found on fore reefs may drive higher rates of dissolved nutrient input \((\text{Atkinson \\& Falter 2003})\). The more complex flows on the fore reef, which includes an along-reef component, suggest caution may be necessary when drawing conclusions from our estimates of fore reef uptake. Nonetheless, our data suggest that the higher fluxes on the fore reef may simply be due to greater concentrations of phytoplankton in this region, which is consistent with theories that the fore reef is the most biogeochemically active zone of the reef and where most reef formation occurs \((\text{Atkinson \\& Falter 2003, Nakamura \\& Nakamori 2007})\).
The observed uptake coefficients $S$ over the reef (~9 to 21 m d$^{-1}$) are comparable to those reported for dissolved nutrients on reefs, 2 to 15 m d$^{-1}$ for NH$_4^+$ and NO$_3^-$ (Atkinson & Falter 2003), possibly suggesting the uptake was limited by hydrodynamic delivery processes (i.e. mass transfer limited). Similar uptake coefficients have been observed in flume studies, with mean (±SD) $S$ values ranging from around 5.4 ± 1 m d$^{-1}$ (Ribes et al. 2003) for a range of particle sizes to 12.9 ± 6.3 m d$^{-1}$ for picoplankton (<2 µm) (Ribes & Atkinson 2007). In flume experiments using both dissolved and particulate materials, $S$ has been observed to be functionally dependent on the local current speed $U$, such that the dimensionless Stanton number ($St = S/U$) is relatively constant between different flow environments (e.g. Baird & Atkinson 1997). Using $St$ thus accounts for flow-dependent effects on $S$ and therefore allows for more direct comparison between studies. Values for $St$ can be calculated based on $m$ and $S$ and an average current velocity on the reef flat of $U = 0.15$ m s$^{-1}$, giving values on the order of $10^{-3}$ (Table 3). This is consistent with the $10^{-3}$ to $10^{-4}$ reported for dissolved nutrient uptake on reefs (Atkinson & Falter 2003) and further suggests that the removal of phytoplankton may be occurring very near to the maximum possible mass transfer limited rate.

Some studies have shown that particle removal rates can actually vary significantly according to particle type, but it is unclear whether this is due to a physical control based on particle size, or a biological control due to preferential grazing by benthic organisms. As detailed above, $S$ values from the flume work of Ribes et al. (2003) suggest that particle uptake is generally higher for smaller particles, with some evidence of preferential removal (e.g. picoplankton over smaller bacteria or larger nano- and microplankton and detrital particles, Ribes et al. 2003; Synechococcus sp. over heterotrophic bacteria or picoeukaryotes, Ribes & Atkinson 2007). Houlbrèque et al. (2006) were the first to demonstrate that preferential removal of picoplankton can also be observed in the field. Our uptake coefficients (~9 to 21 m d$^{-1}$) were most comparable to those previously determined for picoplankton (<2 µm), 12.9 ± 6.3 m d$^{-1}$ (Ribes & Atkinson 2007), suggesting that the bulk of uptake may have been of smaller cells. However, our analysis of size-based phytoplankton uptake was at a lower resolution than that of the previous studies, comparing only large (>5 µm) and small cells. Although there was a significant seasonal difference in the ratio of small to large cells, and there appeared to be fewer small cells on average within the reef (especially in May 2008), cross-reef declines in the ratio of small to large cells were not significant—perhaps a reflection of the high variability inherent in ratio measurements. However, smaller size fractions are recognised to be the most abundant particles in oligotrophic waters (Ducklow 1990), e.g. Ribes et al. (2003) observed that 92% of particulate N in a flume was composed of picoplankton (<2 µm). The greater quantity of small cells supplied to the reef, around 83% <5 µm (Stns 1 and 2, Table 3), and the fact that the ratio of small to large cells did not increase across the reef, confirms greater uptake occurred in the smaller fractions. Indeed, preliminary evidence from much higher resolution analysis confirms that picoplankton abundance (Prochlorococcus, Synechococcus and larger picoeukaryotes) declined across the reef concomitant with bulk phytoplankton uptake (N. Patten et al. unpubl. data). Even in the absence of preferential uptake, the high abundance of small cells places them as an abundant source in a low nutrient environment. Given the apparent role of picoplankton in reef trophodynamics (see also Yahel et al. 1998, Houlbrèque et al. 2004b, 2006), potential physical and biological controls on preferential uptake of smaller particles in the field would be worthy of more detailed examination.

We finally note that the fate of phytoplankton-deplete lagoon water exiting the channel to the reef slope could ultimately play some role in the overall cross-shelf exchange of phytoplankton between the ocean and reef. Visual observations of fronts adjacent to channels suggest that channel outflow is rapidly mixed with offshore water within a few hundred meters. While the reef slope water at Stn 2 could actually contain some unknown (albeit small) fraction of recirculated lagoon water, the much greater volume of water on the slope relative to the lagoon (due to much greater depth) suggests that lagoon dilution should have a minimal influence on the biogeochemistry of slope water, except directly adjacent to channels. Indeed preliminary analysis of the biogeochemical properties of water at Stn 2 suggest a predominately oceanic origin; POC isotopes ($\delta^{13}$C) and C:N ratios increase across the reef and are significantly higher in the lagoon than at Stn 2, which in turn is not significantly different to offshore values (i.e. Stn 1, A. S. J. Wyatt et al. unpubl. data). This argues against a biogeochemically significant influence from channel outflows on slope waters, at least at the distance from channel examined in the present study (~900 m).

**Implications for reef nutrient budgets**

The bulk phytoplankton N flux (all sizes) documented is comparable to typical uptake rates observed for dissolved N (Atkinson & Falter 2003) and confirms that particle feeding may supply the N ‘missing’ in reef N budgets (Atkinson & Falter 2003). Our estimates of phytoplankton PON flux, 2 to 5 mmol N m$^{-2}$ d$^{-1}$ (Table 3),
are among the first direct observations of phytoplankton-driven N flux. Fabricius & Dommasse (2000) found that total suspended particulate matter (SPM) contributed 7 to 15 mmol N m\(^{-2}\) d\(^{-1}\) to a soft coral-dominated reef. Our results suggest that phytoplankton alone may play a significant role in N flux, in addition to that already acknowledged for bulk SPM. In a recent study, Genin et al. (2009) indirectly calculated the N flux due to phytoplankton in the Red Sea to be on average 2.8 mmol N m\(^{-2}\) d\(^{-1}\), based on chl \(\alpha\) and assumed phytoplankton C:N ratios. This estimate lies within the range of PON\(_p\) flux we have observed. Given the high N flux from phytoplankton alone, it is clear that particulate material is indeed likely to be a significant component of N budgets at the reef scale (Atkinson & Falter 2003).

Uptake rates for other sources of N are yet to be quantified for Ningaloo Reef, but some preliminary estimates allow comparison to the PON\(_p\) observed. Nitrate and ammonium concentrations averaged 0.55 ± 0.05 and 0.15 ± 0.01 mmol m\(^{-3}\), respectively, in May 2007 (authors’ unpubl. data). Using these concentrations and a typical S value of 15 m d\(^{-1}\) (Atkinson & Falter 2003), the estimated dissolved N uptake for our study site would be 8 and 2.25 mmol N m\(^{-2}\) d\(^{-1}\) for nitrate and ammonia, respectively. The phytoplankton flux of ~2 to 5 mmol N m\(^{-2}\) d\(^{-1}\) (Table 3) is thus clearly a significant source of N relative to dissolved species of N. The average C:N ratio for typical phytoplankton feeders on this section of Ningaloo is ~8 (for sponges 7.9 ± 0.8, ascidians 8.6 ± 0.4 and soft coral 8.3 ± 0.5, A. S. J. Wyatt et al. unpubl. data). Based on Eqs. (3) & (4), the C:N ratio of phytoplankton supplied to the reef can be estimated as ~4.9. Phytoplankton are thus a rich N source for these benthic communities.

Phytoplankton flux may also be significant for supplying C to reef communities. Most reefs are largely autotrophic, including the Ningaloo reef flat (Beaton 2008, J. L. Falter et al. unpubl. data), and particulate C fluxes have typically been considered insignificant. However, given that large amounts of C are produced and potentially exported from reef systems in organic and inorganic form, such as mucus and carbonate sand, particulate C import deserves further consideration. Integrated net community productivity on the Sandy Bay reef flat was observed to range from 15 to 140 mmol C m\(^{-2}\) d\(^{-1}\) (average 75.8 ± 24) over 4 d in November 2007 (Beaton 2008, J. L. Falter et al. unpubl. data). The phytoplankton C flux on the reef flat of ~20 mmol C m\(^{-2}\) d\(^{-1}\) is thus significant relative to net community metabolism. Two previous studies have also estimated particulate flux to be on the same order as net community productivity (Ayukai 1995, Fabricius & Dommasse 2000), although both studies estimated POC\(_p\) indirectly based on chl \(\alpha\) C conversion ratios and the latter focused on bulk SPM uptake (including detritus as well as phytoplankton) for a soft coral reef. Fabricius & Dommasse (2000) suggested that phytoplankton C flux on soft coral reefs is generally an order of magnitude higher than on hard coral reefs: 125 to 720 g C m\(^{-2}\) yr\(^{-1}\) based on 3 soft coral studies, compared to ~32 to 36 g C m\(^{-2}\) yr\(^{-1}\) over hard coral (Glynn 1973, Ayukai 1995). The hard coral studies cited may have underestimated phytoplankton C flux, with Glynn (1973) neglecting the most abundant smaller size fractions and Ayukai (1995) using a published C:chl \(\alpha\) ratio of 30 to calculate POC\(_p\) (our data suggested a higher value of ~42 is more appropriate). Thus reef flat POC\(_p\) uptake reported here, equivalent to ~88 g C m\(^{-2}\) yr\(^{-1}\), is higher than previously documented for a hard coral reef and confirms the potential importance of phytoplankton C flux in these systems.

**Temporal scales of phytoplankton supply**

Significant uptake of phytoplankton by the reef implies a connection with offshore production and processes occurring at a range of temporal scales, from daily to seasonal. At a seasonal scale, changes in phytoplankton supply to Ningaloo were expected a priori based on knowledge of the oceanographic changes that occur adjacent to the reef during the upwelling season (Hanson et al. 2005, Woo et al. 2006a). The only previous study of upwelling-induced changes adjacent to Ningaloo (Hanson et al. 2005) suggested that increased phytoplankton biomass and productivity would be expected as a result of the upwelling increasing nitrate concentrations. However, Hanson et al. (2005) was based on comparison of LC and NC waters during the upwelling season only (November 2000). The present study is the first to explicitly examine seasonal changes in biological oceanography adjacent to the reef and suggests that phytoplankton supply may actually be highest during the non-upwelling season. Mean (±SE) phytoplankton concentrations adjacent to the reef were 0.7 ± 0.06 during May compared to 0.4 ± 0.04 µg l\(^{-1}\) during November (Fig. 4). Correspondingly, the integrated flux of phytoplankton to the reef during November 2008 (2900 mg m\(^{-2}\) d\(^{-1}\)) was approximately half that observed in May 2007 and 2008 (5000 and 7600 mg m\(^{-2}\) d\(^{-1}\), respectively).

There are several possible mechanisms for increased phytoplankton biomass during the non-upwelling season in May, including (1) autumn acceleration of the LC transporting nutrients from the north, from below the photic zone, and in eddies and meanders; (2) convective mixing due to cooling and storms activity (Koslow et al. 2008); and (3) changes in near-reef biogeochemistry related to autumn coral spawning (i.e. high concentrations of nutrients). Our observations in
May 2007 and 2008 were not preceded by any significant increases in storm activity or cooling that would explain higher phytoplankton productivity. Winter storms may be more of a factor in productivity of the LC off southwest Western Australia (e.g. Koslow et al. 2008). Increased pelagic productivity adjacent to the reef in autumn is possible due to increased nutrient availability following mass spawning (e.g. Wild et al. 2008). This occurs in March to April each year (Simpson 1991) and has been implicated in phenomena such as the aggregation of whale sharks *Rhincodon typus* at the reef from April to May (Taylor 1996, Taylor & Pearce 1999). Examination of satellite images shows that surface expression of increased phytoplankton in the autumn occurred over a vast (at least 100s of km²) spatial scale, and corresponded well with increasing LC flow, placing reef-derived nutrients as an unlikely factor in the autumn bloom and firmly implicating the LC flow, placing reef-derived nutrients as an unlikely factor in the autumn bloom and firmly implicating the broad-scale influence of the accelerating LC in reef-level nutrient fluxes as well as, perhaps, the seasonal dynamics of megafauna such as the whale shark. Interestingly, increased whale shark abundance has also been linked to years of stronger LC flow and phytoplankton biomass, not to upwelling (Wilson et al. 2001, Sleeman et al. 2007).

Upwelling at Ningaloo is considered to be sporadic (Hanson et al. 2005) and requires a significant period of strong southerly winds (~3 d in excess of 5 m s⁻¹) to develop (Willis 2007). It is possible that we did not observe increased phytoplankton concentrations during the upwelling season due to a lack of favourable winds during our November sampling. However, the daily-scale variability in phytoplankton in November was indicative of upwelling. The changes shown in Fig. 6a were also visually evident during diving operations on the fore reef: poor visibility conditions (<5 m) would alternate with exceptional visibility (≥200 m) over several days. It may be that phytoplankton concentrations were only lower in November on average; while the average total flux of phytoplankton to the reef (M) was lower in November 2008, on some days the flux was as large as the average seen in May (~8000 mg chl a m⁻¹ d⁻¹, Fig. 8).

Two principal mechanisms are considered possible in driving the daily-scale variability observed during the upwelling period: changes in the balance between the LC and NC and internal wave activity. The former is best supported by the present study. Changes in wind stress mean that the LC may pulse onto the shelf when the NC weakens, bringing with it water with lower nutrient and phytoplankton concentrations than present in the NC at this time of year. Pulsing of LC water onto the shelf is supported by the shift to southerly flowing (Fig. 6b), warmer (Fig. 6c) water indicative of the LC during periods of reduced phytoplankton concentrations (Fig. 6a). The potential role of internal wave activity in temporally variable nutrient supply requires further consideration, especially at shorter time scales than considered in the present study.

### Spatial scales of phytoplankton supply

Given apparently strong external forcing at a variety of temporal scales, it becomes critical to consider the spatial scale over which oceanographic processes may affect reef trophodynamics. While seasonal dynamics in pelagic production off Ningaloo have yet to be adequately resolved, the present study and that of Hanson et al. (2005) demonstrate the potential for Ningaloo to interact with distinctly seasonal regional oceanographic processes. This suggests the potential for mechanistic linkages between the reef and ocean over a vast spatial scale.

With this variable oceanographic backdrop in mind, we can estimate the scale at which changes in phytoplankton biomass and production would likely influence supply to Ningaloo Reef. We postulate the idea of a dynamic oceanographic ‘catchment’ for reefs, which will vary seasonally depending on both reef-level processes and variability in regional oceanography (Fig. 9). Although analogous to a river catchment, which has fixed spatial dimensions, the oceanographic catchment must be viewed as dynamic, with dimensions that can vary temporarily dependent on offshore production and hydrodynamics.

Uptake along Ningaloo occurs over a length fixed by the length of the reef (L) at a rate M (the spatially integrated phytoplankton uptake rate per unit of along-shore reef width, mg chl a m⁻¹ d⁻¹, Eq. 2). The total phytoplankton flux to the reef (ML) must be balanced by the available phytoplankton per unit time in offshore waters (*A*c), where *c* is the depth-integrated chl a concentration. This balance gives:

\[
A_c = \frac{ML}{c}
\]

where *A*c is the catchment area per day. If we assume that phytoplankton uptake occurs along the length of Ningaloo at similar rates (i.e. Sandy Bay M is representative) and that 85% of Ningaloo is composed of reef with similar uptake area as that at Sandy Bay (measurements from aerial photographs indicate ~87% reef for the ~30 km surrounding Sandy Bay), this conservatively gives an active uptake length (L*) of 246 km for the whole reef. Using the average M observed in May and November 2008, we obtain a phytoplankton flux (ML) of 19 × 10⁸ and 7 × 10⁸ mg d⁻¹ for the LC and NC seasons, respectively. Assuming that depth-integrated chl a concentrations (*c*) are average at non-upwelling
plankton in the water column over 7000 to 30 000 km². Reef communities could completely consume all phytoplankton (Table 4). This implies that in 1 yr, uptake by Ningaloo and NC water is ~87 and ~20 km² d⁻¹, respectively. The plankton flux to the reef must be balanced by net production in the catchment: (i.e. in the absence of significant LC and NC advection). If we assume that net production is the dominant source replenishing phytoplankton in the catchment (i.e. in the absence of significant LC and NC advection), then the plankton flux to the reef must be balanced by net production in the catchment:

\[(P - G)A_c = ML_r\]  

So for the catchment sizes calculated with Eq. (5) (Table 4), this implies that cumulative daily net production would have to be equal to depth-integrated chlorophyll concentrations \(\bar{c}\). Assuming that average gross production \(P\) values observed by Hanson et al. (2005) in LC and NC waters are representative of non-upwelling and upwelling rates (converting from mg C m⁻² d⁻¹ to mg chl a m⁻² d⁻¹ using Eq. 4), and that pelagic grazing rates \(G\) are negligible, it is clear that production is insufficient to maintain the catchments’ phytoplankton concentrations. Gross production in the LC of 4 mg chl a m⁻² d⁻¹ represents ~19% of the phytoplankton required each day by the reef, while in the NC production of 25 mg chl a m⁻² d⁻¹ can supply ~70% of the requirement. So the role of productivity in supporting the reef will depend on the dynamics of the LC and NC, and may be particularly significant for the NC.

Alternatively, the advective flux required to maintain the catchment can be calculated based on the deficit between \(\bar{c}\) and net production (assumed to be \(P\)). Thus ~17.4 and 10.9 mg chl a m⁻² d⁻¹ is required from the LC and NC, respectively. For the catchment areas in Table 4, this equates to ~15 × 10⁸ and 2 × 10⁸ mg d⁻¹. Based on the maximum volume transport \((Q)\) in the LC and NC, ~7 Sverdrups (Sv) (Godfrey & Ridgway 1985, Pearce 1991) and ~0.1 Sv (Woo et al. 2006b), respectively, we can estimate the peak advective fluxes within the 2 currents, \(F_{LC}\) and \(F_{NC}\), using average depth-integrated phytoplankton biomass \(\bar{c}\) as per Table 4 and integration depth \(d\) (~100 and 80 m, respectively, Hanson et al. 2005):

\[F = \frac{Q\bar{c}}{d}\]  

Table 4. Ocean catchment \((A_c, \text{km}^2 \text{d}^{-1})\) for Ningaloo Reef for a catchment dominated by the Leeuwin (LC) and Ningaloo (NC) current. Calculations use depth-integrated phytoplankton biomass \(\bar{c}\) (mg chl a m⁻²) and average reef uptake rates \(ML_r\) (×10⁸ mg chl a d⁻¹). The gross production \(P\) (converted to mg chl a m⁻² d⁻¹ using Eq. 4) is insufficient to maintain \(\bar{c}\) in the catchment. The maximum advective flux in the LC and NC is more than adequate to maintain \(\bar{c}\), with the proportion required in addition to \(P\) shown (%\(F\)). The proportion of advective flux required is likely to have been underestimated since no loss of phytoplankton due to pelagic grazing \(G\) (mg chl a m⁻² d⁻¹) is included in the calculation as it is currently not well quantified for the region.

<table>
<thead>
<tr>
<th>Water mass</th>
<th>(\bar{c}^a)</th>
<th>(ML_r)</th>
<th>(A_c)</th>
<th>(P^a)</th>
<th>%(F^b)</th>
</tr>
</thead>
<tbody>
<tr>
<td>LC</td>
<td>21.4</td>
<td>19</td>
<td>87</td>
<td>4</td>
<td>1.2</td>
</tr>
<tr>
<td>NC</td>
<td>35.9</td>
<td>7</td>
<td>20</td>
<td>25</td>
<td>7.0</td>
</tr>
</tbody>
</table>

*aMeans in vicinity of Ningaloo Reef Hanson et al. (2005)

*bBased on maximum volume transport Woo et al. (2006b)
Rodriguez et al. (2008) observed that C:N ratios in the phytoplankton into the catchment adjacent to Ningaloo is essential to maintain existing reef uptake rates. These first estimates of a dynamic ocean catchment suggest that Ningaloo may have ecological connectivity with the surrounding ocean at scales hitherto unconsidered, and provide a mechanistic basis for examining processes that alter the functional scale of reefs, including changes in regional currents and net productivity.

Implications of the ocean catchment

A dependence of reefs on large-scale, dynamic oceanic catchments has several implications. Firstly, reefs may be even more susceptible to global changes that alter regional oceanography than presently thought. It is recognised that the direct effects of global climate change and ocean acidification may lead to the loss of coral-dominated ecosystems within this century (Hoegh-Guldberg et al. 2007). The role of indirect effects, such as changes in offshore currents and planktonic communities, has not previously been considered. For instance, oceanographic conditions offshore from Ningaloo are driven by climatic conditions that vary seasonally and interannually (Feng et al. 2003). How climate change will alter these conditions, and therefore the production and supply of oceanic material to the reef, is currently unknown. While corals themselves have often been thought to be one of the most sensitive indicators of climate change, planktonic communities may be equally, if not more, sensitive to global change, especially given their short life cycles (Hays et al. 2005, Hooff & Peterson 2006). Climate change and acidification has the potential to alter both the composition of planktonic communities (e.g. Hooff & Peterson 2006) and, perhaps more importantly, their food source. For example, an increase in phytoplankton C:N ratios in response to rising CO2 concentrations (e.g. Riebesell et al. 2007, Iglesias-Rodriguez et al. 2008) would reduce the quality of particulate nutrients supplied to reefs. Iglesias-Rodriguez et al. (2008) observed that C:N ratios in the coccolithophore Emiliania huxleyi increased significantly, from 6.8 to 8.3, when CO2 concentrations increased between 280 and 750 parts per million by volume. If similar adjustment in C:N ratios occurred community wide, and without a concomitant increase in production, the phytoplankton N flux described in Table 3 would likely be altered; foreseeably, 22% more plankton would need to be supplied to the reef for the same N input. These potential indirect effects of climate change and ocean acidification on reef systems require further consideration in order to fully understand the future of coral reefs under various climate change scenarios.

The second implication of the dynamic ocean catchment is its role in sustaining reef biodiversity by supporting reef biota that ultimately depend on ocean supply. The organisms that link reef and ocean through particle uptake may have a disproportionately high significance to a reefs’ function, as well as potentially being very susceptible to changes in the surrounding ocean. Likely candidates for particle dependence include cryptic suspension feeders, such as sponges and ascidians (Richter & Wunsch 1999, Ribes et al. 2005, de Goeij & van Duyl 2007), and planktivorous fishes (Hamner et al. 1988, Pinnegar & Polunin 2006, Hamner et al. 2007). Interestingly, visual observations at our site indicate planktivorous fishes concentrate against the reef wall (Stns 2 to 3, Fig. 4a), most noticeably large schools of the dark-banded fusilier Pterocaesio tile and various pomacentrids, perhaps creating the ‘wall of mouths’ of Hamner et al. (1988). Further work is required to identify which components of the food web play a significant role in the incorporation of particles from the ocean and their susceptibility to oceanographic changes.

Finally, the concept of a dynamic oceanic catchment has broader implications for a range of marine systems and problems, including marine reserve design and the effects of global climate change. Marine reserves designed to protect benthic marine systems may be grossly inadequate if the system is dependent on a large-scale oceanographic catchment that remains unprotected. Similarly, dependence on large-scale oceanographic processes may enhance or ameliorate the effects of global climate change on benthic systems of conservation significance, such as through regional shifts in currents and offshore production. Understanding seasonal changes in an ocean catchment offers the potential to manage the timing of detrimental activities, such as dredging and drilling, in order to minimise the likelihood of negative effects on the catchment and associated reefs. Coral reef biogeochemistry, and many other marine fields, would likely be enhanced by quantifying the role of broad-scale oceanography in local processes, viewing the benthos and surrounding oceanographic regimes as a unified system.

Acknowledgements. We thank J. Erez of The Hebrew University, Jerusalem, for early discussion with A.M.W. We are especially grateful for discussions with J. Falter and M. Atkinson of the Hawaii Institute of Marine Biology over the course of this project. D. Krikke, F. McGregor, S. Hinrichs, N. Patten, A. Chalmers, K. Brooks, D. Kappeli and K. Philip provided valuable sampling assistance. We also thank W. Klonowski and M. Lynch of Curtin University for providing access to processed hyperspectral bathymetry data. Funding was provided by a Natural Environment Research Council Advanced Fellowship and a grant from the British Ecological Society to S.H., grants from The University of Western Australia, the Faculty of Engineering, Computing and Mathematical Sciences, and the Western Australian Marine Science Institution (Node 3) to A.M.W., an Australian Research Council (ARC) Discovery
LITERATURE CITED


► Hatcher BG (1997) Coral reef ecosystems: How much greater is the whole than the sum of the parts? Coral Reefs 16: 587–591


Editorial responsibility: Alain Vézina, Dartmouth, Nova Scotia, Canada

Proofs received from author(s): April 9, 2010

Submitted: May 22, 2009; Accepted: January 24, 2010